

# RESEARCH ARTICLE

# Diverse and highly active diazotrophic assemblages inhabit ephemerally wetted soils of the Antarctic Dry Valleys

Thomas D. Niederberger<sup>1</sup>, Jill A. Sohm<sup>2</sup>, Joëlle Tirindelli<sup>3</sup>, Troy Gunderson<sup>2</sup>, Douglas G. Capone<sup>2</sup>, Edward J. Carpenter<sup>3</sup> & Stephen C. Cary<sup>1,4</sup>

<sup>1</sup>College of Marine and Earth Sciences, University of Delaware, Lewes, DE, USA; <sup>2</sup>Wrigley Institute for Environmental Studies and Department of Biological Sciences, University of Southern California, Los Angeles, CA, USA; <sup>3</sup>Romberg Tiburon Center, San Francisco State University, Tiburon, CA, USA; and <sup>4</sup>Department of Biological Sciences, University of Waikato, Hamilton, New Zealand

Correspondence: Stephen C. Cary, College of Earth, Ocean and Environment, University of Delaware, 700 Pilottown Road, Lewes, DE 19958, USA. Tel.: +1 302 645 4288; fax: +1 302 645 4007; e-mail: caryc@udel.edu

Received 5 October 2011; revised 4 April 2012; accepted 10 April 2012. Final version published online 21 May 2012.

DOI: 10.1111/j.1574-6941.2012.01390.x

Editor: Dirk Wagner

#### Keywords

diazotrophs; Dry Valley; Antarctica; activity; community.

#### **Abstract**

Eolian transport of biomass from ephemerally wetted soils, associated with summer glacial meltwater runoffs and lake edges, to low-productivity areas of the Antarctic Dry Valleys (DV) has been postulated to be an important source of organic matter (fixed nitrogen and fixed carbon) to the entire DV ecosystem. However, descriptions and identification of the microbial members responsible for N<sub>2</sub> fixation within these wetted sites are limited. In this study, N<sub>2</sub> fixers from wetted soils were identified by direct nifH gene sequencing and their in situ N<sub>2</sub> fixation activities documented via acetylene reduction and RNA-based quantitative PCR assays. Shannon-index nifH diversity levels ranged between 1.8 and 2.6 and included the expected cyanobacterial signatures and a large number of phylotypes related to the gamma-, beta-, alpha-, and delta-proteobacteria. N2 fixation rates ranged between approximately 0.5 and 6 nmol N cm<sup>-3</sup> h<sup>-1</sup> with measurements indicating that approximately 50% of this activity was linked with sulfate reduction at some sites. Comparisons with proximal dry soils also suggested that these communities are not ubiquitously distributed, and conditions unrelated to moisture content may define the composition, diversity, or habitat suitability of the microbial communities within wetted soils of the DVs.

# **Introduction**

The McMurdo Dry Valleys (DV) of Antarctica comprise the largest (approximately 15 000 km²) perpetually icefree regions in the Antarctic continent (Hopkins *et al.*, 2006b; McKnight *et al.*, 1999). A combination of prevailing katabatic winds and extremely low precipitation rates renders the DV arguably one of the coldest, driest deserts on Earth (Barrett *et al.*, 2002; Hopkins *et al.*, 2006b; Cary *et al.*, 2010). The valleys were originally formed by successive glaciation events with soils derived from glacial tills and weathered bedrock (Cary *et al.*, 2010). DV soils are characterized by a complete lack of vascular plants, high pH, low organic matter, high salt levels, and mass water content typically below 2% (Virginia & Wall, 1999; Hopkins *et al.*, 2006b; Barrett *et al.*, 2007; Cary *et al.*, 2010).

Biogeochemical cycling of nitrogen (N) and primary sources of N in the bulk arid, severely oligotrophic soils of the DV are still poorly understood (Barrett *et al.*,

2007; Hopkins et al., 2008; Cary et al., 2010). It has been hypothesized that the eolian transportation of microbial biomass from high-productivity zones to low-productivity bulk arid soils may be a major dispersal mechanism of organic matter and therefore an important source of fixed N to the entire DV ecosystem (Gordon et al., 1998; Hopkins et al., 2006a, b; Barrett et al., 2007; Novis et al., 2007; Wood et al., 2008; Pointing et al., 2009; Cary et al., 2010). High-productivity sites in the DV soils are characterized as biological crust communities that form in ephemerally wet soils associated with summer glacial meltwater streams and lake edges. During the short Antarctic summer, temperatures become warm enough for melting to occur on the exposed surfaces of glaciers, resulting in the formation of short-lived (approximately 4 −10 weeks) meltwater streams. These streams are the hydrological link between glaciers and lakes and represent the main source of liquid water, ions, and nutrients to the DV lakes and surrounding soils (McKnight et al.,

2004). Microbial communities that form in the wet soils adjacent to the source (lake/stream) are typically cyanobacterial and/or moss dominated, exist as a thin crust-like layer binding the top 1-5 cm of soil together and have a patchy distribution (McKnight et al., 2004; Adams et al., 2006; Takacs-Vesbach et al., 2010). High flow rates and/ or the steep reaches of stream runoffs may limit the development of these communities owing to the associated turbulences and abrasions of the hydrological flow and are therefore restricted to the zones adjacent to stream banks and lake edges (McKnight et al., 1999, 2007). These communities survive the winter months in a desiccated state and in some cases are re-activated through hydration by summer meltwaters (Takacs-Vesbach et al., 2010), and as a result, they can form large concentrations of biomass even under the extreme in situ environmental conditions (Vincent & Howard-Williams, 1986; McKnight et al., 1999).

Ephemerally wet soils of the DV are documented as hotspots of microbial activity; however, previous studies have focused on measurements of bulk geochemical processes as opposed to identifying the microbial members involved in these processes (Vincent, 1988; Runkel  $et\ al.$ , 1998; McKnight  $et\ al.$ , 1999, 2004, 2007; Virginia & Wall, 1999; Maurice  $et\ al.$ , 2002; Gooseff  $et\ al.$ , 2004; Hopkins  $et\ al.$ , 2006a; Takacs-Vesbach  $et\ al.$ , 2010). From classical morphological-based observations, cyanobacteria have been postulated to be the dominant primary producer and  $N_2$  fixer, with black biomass dominated by *Nostoc* sp., green-colored by *Prasiola* sp. and orange/red by *Oscillatoria* and *Phormidium* (Vincent & Howard-Williams, 1986; Vincent  $et\ al.$ , 1993; McKnight  $et\ al.$ , 1999; Vincent, 2000; Takacs-Vesbach  $et\ al.$ , 2010).

Although cyanobacteria in wetted soils of the DV may represent keystone species, indirectly supplying fixed N to the entire DV ecosystem, comprehensive characterization via modern molecular methods and their contribution to total N budgets remain unresolved. The presence and N inputs by nonphototrophic (i.e. heterotrophic) N fixers in DV soils are also currently unknown. Heterotrophicassociated N2 fixation has been shown to account for a large contribution (approximately 10-30%) of N<sub>2</sub> fixed in Lake Bonney ice aggregates in the DV (Olson et al., 1998; Paerl & Priscu, 1998), and nonphototrophic diazotrophs have been detected in soils associated with Canada stream in the DV (Olson et al., 1998). Therefore, to address these knowledge gaps and through the complementation of classical activity-based analyses (e.g. acetylene reduction assays) with molecular-based genetic methodology, we strived to resolve the identity, diversity, and activity of both photo- and heterotrophic N fixers in wetted soils of the DV. The nifH gene, encoding a subunit of the iron protein of the nitrogenase reductase complex was used as

a molecular marker for N<sub>2</sub> fixation as it yields tree topologies similar to 16S rRNA gene phylogeny (Young, 1992; Zehr *et al.*, 2003a, b). This work represents the first comprehensive study of N<sub>2</sub> fixation in DV soils and the microbial communities associated with these activities.

#### **Materials and methods**

# Sampling and soil characteristics

Samples for this study were collected from the Miers Valley region of Antarctica (Fig. 1). Miers Valley is located at the southern region of the McMurdo Dry Valleys (DV) in Victoria Land and has two glaciers located at the upper (west) end of the valley: Miers Glacier is situated at the northern side and Adams Glacier on the southern side of the Valley. During the summer months, meltwater streams from both glaciers flow into the permanently icecovered Lake Miers (Fig. 1). A number of hyporheic zones and wetted lake boundaries as presented in Fig. 1 were sampled along transects (as described later) during a 2-week period of January 2009. Four transects are described in detail in this study, including: (1) MS1, Miers Stream 1 (S78°05.748', E163°44.822'), a site located at the edge of the runoff stream flowing from Miers Glacier into Miers Lake in Miers Valley; (2) ML1, Miers Lake 1 (S78°05.615', E163°49.912') located on the north side of Miers Lake; and (3) BL, Baby Buddha Lake (S78°03.621', E163°46.463'), and (4) NP, Nostoc Pond (S78°03.920', E163°46.497') adjacent to lake systems located over the north ridge of Miers Valley. Transects consisted of 4 sampling sites originating from site 1 defined as a 'wet' zone (sampling within the lake/stream edge) extending through moist soil zones containing crust communities (sites 2 and 3) to the final site 4 located in a typical Dry valley desert mineral soil. Soil samples were collected aseptically at each site into sterile Whirl-Pak bags using a sterile metal spatula and kept frozen during transport to the laboratory. A second sampling season (November-December 2009) was also carried out, encompassing nitrogenase activity assays with some transects from the January 2009 expedition being resampled. Soil pH was determined on slurries composed of 2 g of soil and 5 mL of de-ionized H<sub>2</sub>O. Percentage of moisture content of soils was calculated following drying for 24 h at 90 °C. Chlorophyll a was measured via the acidification method of Holm-Hansen et al. (1965). Nitrate, silicate, and orthophosphate were analyzed using the Bran and Luebbe AutoAnalyzer II according to standard Bran and Luebbe AutoAnalyzer Applications procedures, G-172-96, G177-96, and G-175-96. Cell counts were determined by the addition of 9 mL of 2% formalin in 0.2-µm filtered de-ionized water to 1 cm<sup>3</sup> of soil. The sample was vortexed for 20 s, placed

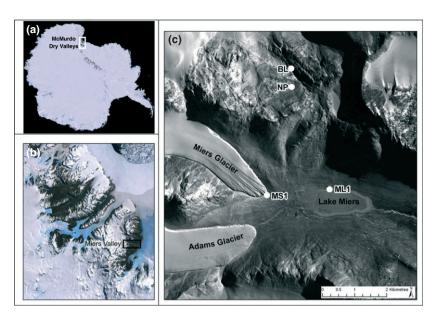


Fig. 1. The McMurdo Dry Valleys and the location of the transect sites in the Miers Valley region. (a) The Antarctic continent with the location of the McMurdo Dry Valleys indicated by the inset box. (b) An enlargement of the a inset box showing the McMurdo Dry Valley region including the Miers Valley study region. (c) An enlargement of Miers Valley and presents the four sampling sites included in the study. Images are reproduced and sourced from the NASA Earth Observatory, the Landsat Image Mosaic of Antarctica (LIMA) Project, and the Advanced Land Observing Satellite (ALOS). This papers incorporates copyright data (JAXA. 2007, Australian Center of Remote Sensing, ACRES) under the permission of the Commonwealth of Australia (Geoscience Australia; 'the Commonwealth') which have not evaluated the data as altered and incorporated within the paper and therefore give no warranty regarding its accuracy, completeness, currency, or suitability for any particular purpose.

in ice, and sonicated for six 20-s 'on' and 20-s 'off' intervals using a Branson 250 sonicator (output set at 5) within a 15-mL polypropylene tube. The sample was allowed to settle for 20 min, and 50  $\mu L$  of a DAPI solution (5 mg mL $^{-1}$ ) was added to 5 mL of the supernatant and incubated for 20 min at room temperature prior to filtering onto a Nuclepore filter (0.2  $\mu m$  pore size; 25 mm diameter). A drop of nonfluorescing immersion oil was added to the filter and a coverslip placed. DAPI fluorescing bacteria were counted at 1000× under oil immersion using a 0.01-mm² grid in the ocular of the microscope.

#### **Nitrogenase activity**

Nitrogenase activity was measured in the field in triplicate using the acetylene ( $C_2H_2$ ) reduction method (Capone, 1993; McKnight *et al.*, 2007). Samples were collected with a plastic syringe with the tip cut off; 41-cm-depth plugs were placed into 27-mL serum vials and sealed. 2 mL of  $C_2H_2$  was added to each bottle, and the increase in ethylene ( $C_2H_4$ ) was monitored in the field for the next 6–10 h on a Shimadzu mini-2 gas chromatograph with a flame ionization detector. Samples were incubated outside in full sunlight. The rate of increase in  $C_2H_4$  is the nitrogenase activity, and this was converted into  $N_2$  fixed with

a conversion factor of  $3C_2H_2/1N_2$  and then multiplied by 2 to account for the two N atoms released per molecule of  $N_2$  fixed. In the second field season (November–December 2009) of this project, we carried out experiments to measure the contribution of  $N_2$  fixation by sulfate reducers; sulfate reduction can be inhibited by  $MoO_4$  when added in concentrations similar to sulfate concentrations found in the surrounding water (Oremland & Capone, 1988). Samples were collected for  $N_2$  fixation as described previously, but a second set of triplicates was inoculated with  $NaMoO_4$  to a final concentration of approximately 1 nM and incubated for 2 h before initiating measurement of  $N_2$  fixation, as described previously.

# **Nucleic acid extractions and cDNA synthesis**

DNA was extracted from samples at the University of Waikato (New Zealand) using the PowerSoil<sup>TM</sup> DNA Isolation kit (MO BIO), as per manufacturer's instructions, and yields quantified using a NanoDrop system (Thermo Fisher Scientific). Total RNA was isolated in the field (a semi-permanent hut in Miers Valley) and at the Crary Laboratory, McMurdo Station, Antarctica, from preserved (LifeGuard<sup>TM</sup>, MO BIO) samples using the RNA Power-Soil<sup>TM</sup> Total RNA Isolation kit (MO BIO). cDNA was prepared at the Crary Laboratory, McMurdo Station,

Antarctica. A total of 16  $\mu$ L of the total RNA extract was added to 2  $\mu$ L of deoxyribonuclease I (1 U/ $\mu$ L; Invitrogen) and 2  $\mu$ L of 10× DNase I reaction buffer (Invitrogen). The reaction was well mixed, incubated at room temperature for 15 min and 2  $\mu$ L of 25 mM EDTA added and incubated at 65 °C for 10 min. cDNA was subsequently constructed from the DNase-treated RNA using the SuperScript III First-Strand Synthesis SuperMix (Invitrogen) with random hexamers, as per manufacturer's instructions. cDNA was stored at -80 °C, and the presence of remnant DNA in RNA extracts and DNase-treated extracts checked via 16S rRNA gene PCR as described later.

# Gene amplification via the polymerase chain reaction (PCR)

All PCR primers utilized in the study are listed in Table 1. Terminal restriction fragment length polymorphism (T-RFLP) PCR was undertaken as described by Danovaro *et al.* (2006). Following PCR, a total of three 25-µL reactions for each DNA sample were combined, and approximately 500–600 ng of purified amplicons digested in duplicate as per manufacturer's instructions using 10 U of *Alu* I (Roche). Duplicate digests were combined and purified as described previously and terminal fragment sizes determined [MegaBACE system (Amersham); Waikato DNA Sequencing Facility, University of Waikato, Hamilton, New Zealand]. Cyanobacterial-specific automated rRNA intergenic spacer analysis (ARISA) PCR was undertaken as described by Wood *et al.* (2008).

Various commonly used *nifH* PCR primer sequences were aligned to multiple *nifH* and alternative dinitrogenase

(anfH and vnfH) gene sequences to check for PCR primer coverage and specificity (results not shown). The most universal degenerated PCR primer pairs were chosen (Table 1), capable of targeting nifH, anfH, and vnfH alternative dinitrogenases using a nested approach based on the protocols of Yeager et al. (2004) and Hewson & Fuhrman (2006). Appropriate negative controls were utilized throughout the molecular analyses, and previously reported contamination problems were not encountered (Zehr et al., 2003a, b). An initial PCR was undertaken utilizing primer pair 19F and nifH3 (Table 1) followed by a nested PCR of the resulting amplicons with primer pair nifH1 and nifH2 (Table 1) resulting in a approximately 370-bp product. The initial PCR comprised of 3 mM MgCl<sub>2</sub>, 1× PCR buffer without MgCl<sub>2</sub> (Invitrogen) 0.2 mM each deoxynucleotide triphosphates, 0.8 µM each primer, 0.5 U Platinum® Taq DNA polymerase (Invitrogen), and approximately 100-200 ng of template DNA or 1 µL DNase-treated RNA extract/cDNA (neat and diluted 1: 10 in pure water) in a final reaction volume of 25 μL. Thermocycling conditions consisted of 95 °C for 5 min followed by 21 cycles of 94 °C for 45 s, 48 °C for 1 min, and 72 °C for 1 min with a final extension step at 72 °C for 10 min. The subsequent nested PCR consisted of 2.5 mM MgCl<sub>2</sub>, 1× PCR buffer without MgCl<sub>2</sub> (100 mM Tris-HCl pH 8.3, 500 mM KCl, SIGMA) 0.2 mM each deoxynucleotide triphosphates, 0.5 µM each primer, 20 µg bovine serum albumin, 1.25 U JumpStart<sup>TM</sup> Tag polymerase (SIGMA), and 1 µL of template (PCR product from the initial PCR) in a final reaction volume of 50 μL. The thermocycling conditions consisted of 94 °C for 5 min followed by 20 cycles of 94 °C for 45 s, 65 °C (−0.5 °C each cycle) for 1 min, and 72 °C for 1 min and then 12 cycles of 94 °C for 45 s, 55 °C for 1 min, and

Table 1. Oligonucleotides utilized in this study

Primer	Purpose	Target (amplicon size)	Sequence (5' -3')	Reference
27F	T-RFLP	Bacteria	FAM-AGA GTT TGA TCC TGG CTC AG	Lane (1991)
1492R			GGT TAC CTT GTT ACG ACT T	
CY-ARISA-F	ARISA	Cyanobacteria	FAM-GYC AYR CCC GAA GTC RTT AC	Wood et al. (2008)
23S30-R			CHT CGC CTC TGT GTG CCW AGG T	
19F	nifH initial amplification	nifH gene	GCI WTY TAY GGI AAR GGI GG	Ueda et al. (1995)
nifH3			ATR TTR TTN GCN GCR TA	Zani et al. (2000)
nifH1	nifH nested		*TGY GAY CCN AAR GCN GA	Zehr & McReynolds
nifH2			ADN GGC ATC ATY TCN CC	(1989)
deltaG1F	Delta-proteobacterial-specific	Group 1 OTUs: 2, 7, 13	AGG CGG ATT CGA CTC GGC TAT	This study
deltaG1R	nifH gene qPCR	and 47 <sup>A</sup> (260 bp)	GCC GCA TAC CAC GTC ACC C	
deltaG2F	Delta-proteobacterial-specific	Group 2: OTUs: 11	CAT GCC AAG GCG CAG AAY ACG	This study
deltaG2R	nifH gene qPCR	and 27 <sup>A</sup> (322 bp)	CCA TCA TYT CGC CGG AGC AG	
cyanoUF	Cyanobacterial-specific nifH	All cyanobacterial	TAT YAT CAC CGC YAT YAA CT	This study
cyanoUR	gene qPCR	OTUs <sup>A</sup> (152 bp)	TAG ATT TCT TGB GCY TTR CC	
betaF	Beta-proteobacterial-specific	OTUs: 1, 10, 31,	GTC ATG MRD ATY GGC TAC AAR	This study
betaR	nifH gene qPCR	33 <sup>A</sup> (182 bp)	GCC TTG TTK TCR CGG ATS GG	

72 °C for 1 min with a final extension step at 72 °C for 10 min.

# **DNA fingerprinting comparisons**

Fluorescent peak data from both the T-RFLP and ARISA assays were uploaded and aligned in the T-REX online platform (Hewson & Fuhrman, 2006). Peaks below 300 bp were manually removed for ARISA as outlined by Wood et al. (2008). Aligned data were imported into PRIMER6 (Primer-E Ltd., Plymouth, UK), and comparison of community fingerprints was undertaken by twodimensional principle component analyses (PCA) with overlaid resemblance matrix similarities, and BEST analysis was used to identify abiotic factors that best explain the observed community fingerprints (Clarke & Gorley, 2006). For BEST analyses, samples ML1-4 and NP4, and cell count data were excluded from the analyses owing to missing data (Table 2). All abiotic variables were log transformed with the expectation of chlorophyll a concentrations (as decided by draftsman plots), normalized, and comparisons made between community structure and abiotic factor similarity matrices (Euclidean distance) using Spearmans rank correlation with 999 permutations.

# Partial length nifH gene cloning

Duplicate nifH1- and nifH2-primed PCR amplicons (as described previously) were combined, purified (GenElute<sup>TM</sup> PCR clean-up kit; SIGMA) and subsequently ligated into the pCR4-TOPO vector (Invitrogen), and transformed into One Shot TOP10 chemically competent *Escherichia coli* cells (Invitrogen) using the TOPO TA cloning kit for Sequencing (Invitrogen). Transformants containing the vector were selected on Luria Broth (LB) with kanamycin (50 µg mL<sup>-1</sup>)-selective solid media, and the resulting plasmids with inserts were Sanger sequenced at the Genome Center at Washington University (St. Louis, MO).

#### Sequence analysis

nifH sequences were aligned using default parameters within CLUSTALW (Thompson et al., 1994) and distances calculated using the Jukes—Cantor correction for multiple substitutions using the DNADIST program of PHYLIP (www.phylip.com). Rarefaction analyses were undertaken using the DOTUR program (Schloss & Handelsma, 2005), and representative nifH sequences (cutoff of 95% sequence similarity) were aligned and translated within the GENEIOUS software environment (www.geneious.com) with respect to translated nifH genes obtained from the NCBI GenBank database. The alignment was manually checked and a phylogenetic tree constructed using

Jukes–Cantor corrected distances and the Neighbor-joining method with 1000 bootstrap re-samplings. *nifH* sequences (cutoff of 95% sequence similarity) have been deposited in the NCBI GenBank database as accession numbers HM140726 to HM140774.

#### Quantitative PCR (qPCR)

qPCR primer pairs specific to the major *nifH* phylogenies (delta-proteobacteria group 1 and 2, cyanobacteria, and beta-proteobacteria, Fig. 4) were identified using the Oligo software package (Version 6; Molecular Biology Insights, Inc., Cascade, CO). Owing to the high dissimilarity between *nifH* sequences within each *nifH* phylogenetic grouping, minor OTUs were removed to deliver primer pairs specific to the major *nifH* clades. Table 1 provides the primer sequences, amplicon size, and the target OTUs. Specificity, including optimization of primer concentrations and annealing temperatures of each primer pair, was determined by screening representative *nifH* clones from each phylogenetic grouping.

The Minimum Information for Publication of Quantitative Real-Time Experiments (MIQE; Bustin et al., 2009) has been followed. qPCR assays were undertaken in 96 well-plate format with an ABI7500 Real-Time PCR system, and results were analyzed using the associated Sequence Detection Software (version 1.2.; Applied Biosystems, ABI, Carlsbad, CA). Triplicate 10-µL assays were undertaken using 5 µL of SYBR® Green PCR master mix (ABI), 1 µL of each primer (9 µM), 10 or 20 ng of template DNA or 1 µL of cDNA diluted 1:10 and 1:50 in PCR grade H<sub>2</sub>O (MO BIO), and the volume brought to 10 μL with PCR grade H<sub>2</sub>O (MO BIO). Thermocycling conditions consisted of 50 °C for 2 min, initial denaturing at 95 °C for 10 min followed by 50 cycles of 95 °C for 15 s, and an annealing and extension step for 1 min at 68 °C for the delta-proteobacterial-specific primer pairs and 60 °C for the cyanobacterial- and beta-proteobacterial-specific primer pairs. All qPCR assays included a final dissociation stage consisting of 95 °C for 15 s, 60 °C for 1 min, and 95 °C for 15 s. Standard curves used to estimate gene copy numbers were determined by the utilization of pCR4-TOPO vectors with inserted nifH as described previously.  $r^2$  values of standard curves were above > 0.98 with slopes ranging between -3.2 and -3.6 equating to efficiencies of 105.35–89.57%. Quantification cycles  $(C_q)$  of the DNA samples ranged between approximately 18-33 and 26-31 for the cDNA samples with Cq values being below detection limits or below 40 for nontemplate controls. Assays were repeated for consistency, and PCR inhibition (i.e. change in  $C_q$  values) of DNA and cDNA amplification was not witnessed when spiked with positive control plasmid solutions.

Table 2. Transect descriptions and associated data. All data collected from initial field season (January 2009)

Location	Transect Iabel	Dist. from Pt 1.	Macroscopic Appearance	N <sub>2</sub> fix. (± S.D.)*	Hd	Moisture content (%)	DNA g <sup>-1</sup> (μg)	Chl. <i>a</i> cc <sup>-1</sup> soil (μg)	Cells in 0-2 cm top soil (cm <sup>-3</sup> )	Cells in porewater (ml <sup>-1</sup> )	Nitrate (NO <sub>3</sub> +NO <sub>2</sub> ) (μM)	Silicate (Si(OH) <sub>4</sub> ) (μM)	Ortho- phosphate (PO <sub>4</sub> ) (μM)
Baby Buddha	BL1	na	Lake edge (saturated black mat)	1.53 (0.75)	7.48	34.14	4.709	0.0526	$1.37 \times 10^{7}$	$3.09 \times 10^4$	0.54	62.39	0.17
Lake	BL2	E (	Yellow-colored mat	2.80 (1.41)	7.66	22.75	2.054	0.1636	$3.65 \times 10^{8}$	$2.28 \times 10^4$	0.43	154.24	89.0
	BL3 BL4	2 m 6 m	Black patchy mat Dry gravel soil	0.04 (0.07)	7.92 9.20	17.92 0.35	1.945 0.036	0.2176 0.2324	$2.76 \times 10^{\circ}$ $3.00 \times 10^{7}$	$1.23 \times 10^{4}$ $3.66 \times 10^{4}$	0.44	162.32 100.17	0.24
Miers	ML1-1	na	Sediment of Lake	0.70 (0.28)	8.59	nd⁺	1.09	0.4857	$1.09 \times 10^{8}$	$8.30 \times 10^{3}$	25.73	63.71	0.79
Lake 1	ML1-2	3 3	Black-colored mat	2.96 (0.39)	7.70	22.73	4.927	0.1020	$9.87 \times 10^{7}$	$3.89 \times 10^{3}$	54.1	130.87	1.87
	ML1-3	3.65 m	Patchy biofilm	0.22 (0.096)	9.52	14.54	1.509	0.2018	$1.21 \times 10^{8}$	$2.69 \times 10^{3}$	17.81	184.78	2.17
	ML1-4	8.3 m	Dry gravel soil	0	9.02	2.1	1.218	0.0104	$2.04 \times 10^{8}$	pu	pu	pu	pu
Miers	MS1-1	na	Stream bed, loose	0	7.01	µpu	lpq	0.0129	pu	$9.21 \times 10^{3}$	44.16	90.07	0.35
Stream 1			gravel										
	MS1-2	9.6 m	Edge of stream bank	1.00 (0.66)	9.26	4.5	2.909	5.0438	$3.57 \times 10^{8}$	$4.46 \times 10^4$	39.42	76.56	0.29
	MS1-3	13.5 m	Extensive dry mat	3.38 (0.88)	8.58	19.2	3.963	0.0598	$8.43 \times 10^{7}$	$3.75 \times 10^4$	36.86	52.41	0.51
	MS1-4	20 m	Dry gravel soil	0.53 (0.33)	9.94	3.2	0.981	0.3284	$5.54 \times 10^{7}$	$3.11 \times 10^4$	22.88	151.35	2.47
Nostoc	NP1	na	Lake edge (red/green	2.03 (0.97)	7.63	32.74	2.627	0.0339	$1.41 \times 10^{9}$	$3.68 \times 10^4$	19.20	170.52	0.42
Pond			mats)										
	NP2	30 cm	Green-black extensive	5.83 (2.11)	7.19	31.35	2.363	0.0608	$9.56 \times 10^{7}$	$1.62 \times 10^4$	25.54	213.82	0.87
			mat										
	NP3	1.8 m	Black moss-like	3.04 (0.17)	8.35	11.43	1.363	0.2586	$1.15 \times 10^{7}$	$5.54 \times 10^4$	29.99	256.34	0.56
			mat/gravel										
	NP4	3.7 m	Dry gravel soil	0	8.86	3.29	0.581	pu	nd	nd	nd	pu	pu

\*units = nmol N cm<sup>-3</sup> h<sup>-1</sup>, with standard deviations.

\*Saturated (sample collected below lake water level). na, not applicable; nd, not determined; bdl, below detection limits.

#### **Results**

# Transects and associated geochemistry

Locations of meltwater-to-dry soil transects sampled are indicated in Fig. 1, and associated geochemical data are presented in Table 2. Microbial communities sampled in the wetted soils existed as a thin (1-2 cm) crust layer and exhibited a patchy distribution. The lengths of the transects varied from 3.7 to 20 m ensuring that a typical arid DV desert soil was always sampled at position 4 in each transect (Table 2). Dry soils (site 4 of each transect) were considerably drier (0.35-3.3%) than that of the remainder of the wetted soils (sites 2, 3; 4.5–31% moisture) and the fully saturated (site 1) locations. Noticeable trends within the transects included a pH shift from neutral (approximately 7.0-8.6) to alkaline (approximately 8.8-9.9) from wet to dry points, and highest concentrations of DNA (approximately 2–5  $\mu g \ g^{-1}$  soil) and  $N_2$  fixation rates (approximately 2.0–5.8 nmol N cm<sup>-3</sup> h<sup>-1</sup>) observed in the wettest soils. Cell counts were higher for wet and dry topsoil (approximately  $10^7 - 10^9$  cells cm<sup>-3</sup>) than that of the associated porewater (approximately  $10^4$  cells mL<sup>-1</sup>). Ranges of nitrate, silicate, and orthophosphate concentrations were also similar between sites from each transect with the exception of lower nitrate concentrations at Baby Buddha Lake (Table 2).

# Microbial community comparisons

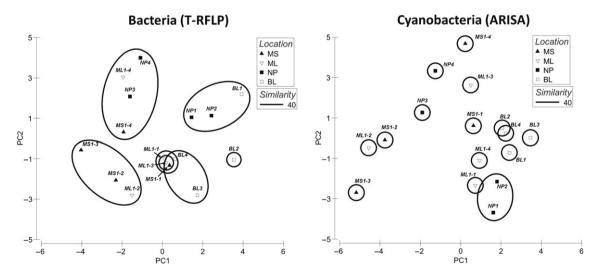
Comparison of bacterial (T-RFLP) and cyanobacterial (ARISA) community fingerprints by principle component analysis is presented in Fig. 2. Distinct clusters at 40%

similarity were apparent for the bacterial fingerprints. Generally, wet sources (site 1) of NP and BL; arid soils of NP, ML1, and MS1; and hyporheic soils (sites 2 and 3) from MS1 and ML1 were grouped together at 40% similarity in the bacterial plot. A discernible distinction in clustering was not observed for the cyanobacterial ARISA-based plot (Fig. 2).

The total number of T-RFs within the T-REX analysis framework for both bacterial T-RFLP and cyanobacterial ARISA analyses were similar between transects and sites within a single transect (Supporting information, Table S1). For example, the average number of T-RFLP and ARISA fragments for the four dry (site 4) soils (61 and 64, respectively, standard deviation approximately 7.0 for both) were similar to the average number of fragments for the wet and moist (sites 1, 2, and 3) transect sites (approximately 56 and 63, standard deviations approximately 13 and 16, respectively). The highest correlation between abiotic measurements (Table 2) and microbial community structure via BEST analyses included distance, N2 fixation rates, and nitrate concentrations ( $\rho = 0.513$ ) for bacterial communities, and distance, N<sub>2</sub> fixation rates, pH, and nitrate concentrations ( $\rho = 0.535$ ) for cyanobacterial communities.

# **Nitrogenase activities**

In situ  $N_2$  fixation rates measured in the first season varied from undetectable to a maximum of approximately 5.8 nmol N cm<sup>-3</sup> h<sup>-1</sup> (Table 2), and as expected, rates were highest from the hyporheic sites (sites 2 and 3) ranging from 0.04 to 5.8 nmol N cm<sup>-3</sup> h<sup>-1</sup> and low or undetectable activities (0–0.5 nmol N cm<sup>-3</sup> h<sup>-1</sup>) in arid

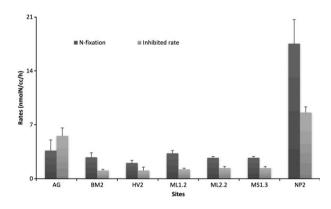


**Fig. 2.** Principle component analyses of bacterial T-RFLP and cyanobacterial ARISA fingerprints. Similarity (%) between fingerprints, as determined by presence/absence of T-RFs, has been overlaid within the PRIMER6 framework.

soils (site 4). The 'wet' source samples (site 1) were also positive for N<sub>2</sub> fixation (0.7–2.0 nmol N cm<sup>-3</sup> h<sup>-1</sup>) with the exception of site 1 of the MS transect, which was collected in the center of the stream where high flow rates presumably inhibited the formation of a stable microbial community. Owing to the discovery of sulfate-reducing N<sub>2</sub> fixers from samples collected during the initial January 2009 sampling season, wetted DV soils were measured for N<sub>2</sub> fixation activities that were linked with sulfate reduction during the second, December 2009 sampling expedition. Through the inhibition of sulfate reduction by the addition of NaMoO4, it was found that sulfate-dependent N<sub>2</sub> fixation activity accounted for a large proportion of the total N2 fixation activities detected within wetted DV sites (Fig. 3). For example, approximately 50% of total N<sub>2</sub> fixation activities for both wetted transect sites, ML1-2 and MS1-3, were linked with sulfate reduction with rates decreasing from approximately 2.7 1.4 nmol N cm<sup>-2</sup> h<sup>-1</sup>.

# nifH gene analyses

The *nifH* gene was not initially detected by PCR in the 'dry' (site 4) soil samples with the exception of BL4, suggesting low gene copy numbers. However, an increase from 21 to 30 PCR cycles during the initial amplification resulted in detectable *nifH* amplicons. The *nifH* gene could not be detected by PCR in samples from the high flow rate Miers Stream (MS1-1) site. *nifH*-based T-RFLP analyses (method by Hewson & Fuhrman, 2006) were undertaken on all transect sites; however, owing to both



**Fig. 3.**  $N_2$  fixation rates measured at various wetted Antarctic soils by the acetylene reduction assay during the second field season (November–December 2009). Activities are presented as both with and without the inhibition of sulfate reduction by the addition of NaMoO<sub>4</sub>. Sites: AG, base of the Adams Glacier; BM, Black Mat from a flooded region of the Miers Valley; HV, site at the shore of the Hidden Valley lake; ML, Miers Lake transects; MS, Miers Stream; NP, Nostoc Pond.

low numbers of *nifH* TRFs (between 2 to 10) and single T-RFs for MS1-4 and ML1-3 (results not shown), results were not analyzed further. *nifH* clone libraries were constructed from the sampling sites with the highest detected N<sub>2</sub> fixation rates from each transect, that is, BL2, NP2, MS1-3, and ML1-2 (Table 3), and the cDNA from ML1-2, as this was the only cDNA sample that provided positive *nifH* PCR amplification (Table 3). Rarefaction curves of *nifH* libraries (Fig. S1) indicate that ML1-2 cDNA has reached its diversity plateau with slopes beginning to plateau for MS1-3 and ML1-2 and steep rarefaction curves for NP2 and BL2, signifying that the diversity of *nifH* in these sites are under-represented. Shannon diversity indices (at 95% sequence similarity) ranged between 1.8 and 2.6 for each individual *nifH* library (Table 3).

Four major nifH clusters were observed (Fig. 4) including two clades (group 1 and 2) within the deltaproteobacteria (one of which grouped closely with Geobacter sp.), one cluster within the cyanobacteria, and a group most closely related to the beta-proteobacteria. nifH phylotypes were also represented within the alphaand gamma-proteobacteria. Table 3 includes the percentages of each phylotype detected within the samples. The MS1-3 and BL2 clone libraries were dominated by deltaproteobacterial-related signatures. Cyanobacterial phylotypes were most well represented in ML1-2, but not detected in BL2. ML1-2 cDNA contained a total of 6 OTUs that were dominated by cyanobacterial-related signatures and a single delta-proteobacterial OTU (OTU23) as opposed to 16 OTUs in DNA from the same site (at 95% sequence similarity). In total, only 3 of the 6 OTUs of the ML1-2 cDNA library were detected in the ML1-2 DNA library. Similarities of nifH phylotypes between sites were typically below approximately 56% with the exception of MS1-3 and ML1-2 sharing approximately 78% of their collective phylotypes (Table S2).

The majority of the *nifH* sequences showed low sequence similarity to *nifH* sequences within the NCBI GenBank database with sequence homologies typically not exceeding 90%. Higher sequence similarities (> 90%) were typically only obtained for *nifH* sequences related to cyanobacterial signatures previously recovered from the ice cover of Lake Bonney, Antarctica (Olson *et al.*, 1998), and those most closely related to clones (OraP15 and OraP17) were recovered from microbial mat communities of a meltwater pool on the McMurdo ice shelf (Jungblut & Neilan, 2010).

Quantitative PCR (qPCR) assays (oligonucleotides listed in Table 1) were developed to target the four major *nifH* phylotypes (delta-proteobacteria group 1 and 2, cyanobacteria and the beta-proteobacteria; Fig. 4) detected in the collective *nifH* clone libraries (Table 4). qPCR targeting these phylotypes were only undertaken on

**Table 3.** Percentage of *nifH* gene phylotypes (as designated in Fig. 4) detected in each clone library. Numbers in parentheses represent the number of clones

A. % nifH clone phylotypes	BL2	NP2	MS1-3	ML1-2	ML1-2 cDNA
Shannon diversity index	2.6	2.1	1.9	2.1	1.8
deltaproteobacteria	75.7 (53)	33.3 (24)	64.3 (45)	43.2 (32)	18.4 (14)
cyanobacteria	nd	13.9 (10)	10 (7)	35.1 (26)	81.6 (62)
gammaproteobacteria	8.6 (6)	4.2 (3)	5.7 (4)	6.8 (5)	nd
alphaproteobacteria	5.7 (4)	nd	nd	12.2 (9)	nd
betaproteobacteria	10 (7)	48.6 (35)	20 (14)	2.7 (2)	nd
Total	100 (70)	100 (72)	100 (70)	100 (74)	100 (76)

the hyporheic soils and indicated that cyanobacterial *nifH* genes were the most abundant (approximately  $10^5$ – $10^7$  copies  $g^{-1}$  soil), with the exception of BL2 (approximately  $10^3$  copies  $g^{-1}$  soil) (Table 4). Delta-proteobacterial phylotypes dominated the BL2 library, approximately  $10^5$  and  $10^4$  copies  $g^{-1}$  soil for groups 1 and 2, respectively (Table 4). Nonphototrophic *nifH* gene copy numbers ranged from approximately  $10^1$  to  $10^5$  for all samples (Table 4). Cyanobacterial *nifH* genes were the only signatures detected in the cDNA fractions (Table 4) from MS1-3 and ML1-2 (approximately  $10^6$  and  $10^7$  copies  $g^{-1}$  soil) with *nifH* being below detection levels for cDNA from both BL2 and NP2.

#### Discussion

The bulk mineral soils of the Antarctic Dry Valleys (DV) are extremely dry (approximately < 2% mass water content), making water availability a particularly important driver of microbial activity (Zeglin et al., 2009; Cary et al., 2010) as is evident from the formation of extensive microbial crust communities associated with or adjacent to perpetually wet areas of the valleys (Kennedy, 1993). While water availability is critical for biomass formation, evidence is accumulating to suggest that other environmental variables may constrain the presence, distribution, and diversity of certain organisms within DV soil habitats. For example, Wood et al. (2008) concluded that moisture content of DV soils was a poor indicator of cvanobacteria distribution and that other variables such as soil elemental composition may play a role in influencing edaphic cyanobacterial habitat suitability. Likewise, Smith et al. (2006) observed a similar trend of cyanobacterial distribution within DV soils. From principal component analysis (PCA) of ARISA-based fingerprints (Fig. 2), our results also suggest that cyanobacterial communities in the DV are not ubiquitously distributed and local conditions unrelated to moisture content may define their composition or habitat suitability. Although cyanobacterial communities differed between biotopes, community structures of both bacteria and cyanobacteria were correlated with distance from the hydrological source, N<sub>2</sub>

fixation rates, pH levels, and nitrate concentrations indicating that, while communities may differ (i.e. no distinction in clustering of ARISA plots in Fig. 2), community structure may be defined by these previously mentioned local factors. Similarly, a recent study by Smith et al. (2010) found high spatial variability of bacterial communities in soils from Taylor Valley in the DV and an icefree coastal region (Cape Hallett) in northern Victoria land that were correlated with soil pH and moisture content ( $\rho = 0.564$ ). In our study, differences in nifH phylotypes between the sampled hydrological sources also suggest a nonubiquitous distribution of microbial communities with wetted DV soils, that is, approximately 80% of nifH phylotypes shared between the hydrologically linked Miers Valley lake (ML1-2) and stream (MS1-3) systems and approximately 50% shared between the lake systems of Nostoc Pond (NP) and Baby Buddha Lake (BL), which were located at higher elevation and in an adjacent valley (Table S2).

As discussed previously, moisture has been hypothesized not to be an important driver of microbial distribution. Studies also suggest that microbial diversity is not influenced by moisture within DV environments (Wood et al., 2008). This has been evidenced by reports of similar Shannon diversity indexes of 16S rRNA gene clone libraries of Antarctic soils of both high (approximately 9% soil moisture) and low (approximately 2-4% soil moisture) primary productivity potential (Niederberger et al., 2008) and similar number (approximately 5-27, depending on biotope) of analysis fragment lengths of cyanobacterial-specific community ARISA profiles between DV lake and soil samples (Wood et al., 2008). We also obtained similar numbers of bacterial and cyanobacterial TRFs from transect sites of contrasting water content (Table S1). This trend has also been noted in the temperate Chihuahuan desert (Liu et al., 2000) and therefore may hold true for wide-ranging desert soil biotopes.

Cell counts in the top 0.2 cm layer of the soil samples were also similar between all soil types tested, that is, approximately  $1.2 \times 10^7$  to  $1.4 \times 10^9$  cm<sup>-3</sup>. These cell density values fall within the cell count ranges previously reported for various DV soil biotopes as estimated by



**Fig. 4.** Phylogenetic tree based on amino acid sequences translated from *nifH* OTUs grouped at 95% DNA sequence similarity for Miers Lake site 2 (ML1-2) DNA and cDNA, Miers Stream site 3 (MS1-3), Nostoc Pond site 2 (NP2), and Buddha Lake site 2 (BL2). The numbers of clones in each group are indicated in parentheses. The tree was constructed from a total of 115 amino acid residues using the Neighbor-joining method with 1000 bootstrap re-samplings. Bootstraps above 50% (i.e. 500 of 1000) are presented. *Methanocaldococcus jannaschii* DSM 2661 (L177117) and *Methanococcus voltae* (X03777) were used as an out-group and have been removed from the tree. The scale bar represents the number of amino acid substitutions per site. Groups 1 and 2 represent the two delta-proteobacterial lineages targeted by quantitative PCR (qPCR).

ATP analyses (Cowan *et al.*, 2002), including, rich/moist soils (1  $\times$  10<sup>7</sup> to 4  $\times$  10<sup>9</sup> g<sup>-1</sup> wet weight) and bulk desert soils (5  $\times$  10<sup>5</sup> to 4  $\times$  10<sup>8</sup> g<sup>-1</sup> wet weight<sup>-1</sup>).

Assuming that the majority of the measured  $N_2$ -fixation rates were in the mats themselves, measured rates (approximately 0.5–6 nmol N cm<sup>-3</sup> h<sup>-1</sup>) can be consid-

ered areal estimates are comparable with areal-based rates detected in stream (0.3 nmol N cm<sup>-2</sup> h<sup>-1</sup>), soil (4.7 nmol N cm<sup>-2</sup> h<sup>-1</sup>), and pond (1.9 nmol N cm<sup>-2</sup> h<sup>-1</sup>) microbial mats of the sub-Antarctic Islands, Livingston Island, and South Shetland Islands (Fernández-Valiente *et al.*, 2007). Similar rates were also previously recorded

**Table 4.** *nifH* gene copy number per gram of soil as determined by gPCR. Numbers in parentheses represent standard deviations of triplicate assays

B. nifH copies $g^{-1}$ soil	BL2	NP2	MS1-3	ML1-2	MS1-3 cDNA	ML1-2 cDNA
delta-proteobacteria G1	$1.20 (0.09) \times 10^{5}$	$3.49 (2.82) \times 10^{1}$	$1.56(0.56) \times 10^2$	lpq	lpq	lpql
delta-proteobacteria G2	$1.29 (0.03) \times 10^4$	$3.43 (1.35) \times 10^{2}$	$9.74 (0.19) \times 10^3$	$2.88 (1.38) \times 10^{2}$	lpq	lpq
cyanobacteria	$3.64 (1.36) \times 10^3$	$2.98 (1.02) \times 10^{6}$	$2.92 (0.23) \times 10^{5}$	$1.90 (0.24) \times 10^{7}$	$2.07 (0.04) \times 10^6$	$2.04 (0.004) \times 10^{7}$
beta-proteobacteria	$5.00 (1.72) \times 10^3$	$4.09 (0.73) \times 10^4$	$3.88(0.97) \times 10^4$	$9.13 (0.92) \times 10^3$	lpq	lpq
nifH copies cm <sup>-3</sup> soil						
delta-proteobacteria G1	$1.74 \times 10^{5}$	$4.85 \times 10^{1}$	$2.81 \times 10^{2}$			
delta-proteobacteria G2	$1.87 \times 10^4$	$4.77 \times 10^{2}$	$1.75 \times 10^4$	$4.58 \times 10^{2}$		
cyanobacteria	$4.93 \times 10^3$	$4.14 \times 10^{6}$	$5.26 \times 10^{5}$	$3.02 \times 10^{7}$	$3.73 \times 10^{6}$	$324 \times 10^{7}$
beta-proteobacteria	$2.25 \times 10^{3}$	$5.69 \times 10^4$	$6.98 \times 10^4$	$1.45 \times 10^4$		

Not detected; bdl, below detection limits.

from Antarctic pond-/lake-associated sites: McMurdo Ice shelf cyanobacterial mats (5.1 nmol N cm<sup>-2</sup> h<sup>-1</sup>; Fernández-Valiente et al., 2001) and lake ice aggregates, for example, 9.4–91.3 nmol N mg<sup>-1</sup> chl a h<sup>-1</sup> (Priscu et al., approximately  $0.08 \text{ nmol N g}^{-1}$  (dry and weight) h<sup>-1</sup> (Olson et al., 1998; Paerl & Priscu, 1998). Interestingly, these rates are of the same magnitude as seen in cyanobacterial mats found in temperate, subtropical, and tropical shallow marine systems such as coral reefs (Charpy et al., 2010) and intertidal lagoons (Steppe & Paerl, 2005). While cyanobacterial mats from warmer climates are composed of different species, the similarity of these rates indicates that temperature may be a driving factor in defining species composition, but not setting the areal N<sub>2</sub> fixation rates possible in these mats.

As typically found in other terrestrial habitats (Izquierdo & Nüsslein, 2006; Zehr et al., 2003a, b), a large diversity of nifH sequences was observed. Shannon diversity index values ranged between 1.8 and 2.6 (Table 3), comparable with soils of the Arctic tundra (1.97), glacier fore-field (2.16-2.46) and tropical forest soil (2.41; Izquierdo & Nüsslein, 2006; Duc et al., 2009). Unexpectedly, cyanobacterial nifH phylotypes did not dominate the gDNA clone libraries; phylotypes from within the cyanobacteria, but also the alpha-, gamma-, delta-proteobacteria, and a clade most closely related to the beta-proteobacteria were all represented. These results mirror results from in marine-associated microbial mats and rice roots, whereby a large diversity of heterotrophic nifH genes have also been found to dominate nifH clone libraries (Ueda et al., 1995; Zehr et al., 1995; Steppe & Paerl, 2002). Low sequence homology (approximately < 90%) of the heterotrophic nifH sequences to their closest NCBI GenBank matches also indicates that they currently remain uncharacterized. Although cyanobacteria did not dominate the gDNA libraries, they were the most abundant phylotype (approximately  $10^5-10^7$  gene copies g<sup>-1</sup> soil) in the wetted soils and were the only phylotypes detected by qPCR in the cDNA fractions for both the Miers Valley transects: ML1-2 and MS1-3. However, it is important to note that the presence of nifH mRNA transcripts may not represent definitive N2 fixation activity as nitrogenase is subject to posttranslational control (Steppe & Paerl, 2002). Also, some cDNA sequences were not retrieved from gDNA clone libraries and may indicate biases in the PCR assays or low copy genes that are highly expressed, whereas gDNA clones not present in cDNA represent genes that were not transcribed at the time of sampling. Likewise, differences between qPCR and cDNA libraries may indicate biases owing to the contrasting PCR conditions, for example, primers and annealing temperatures.

Olson et al. (1998) also documented nifH sequences from Antarctic mat, soil, and ice samples, including Lake Bonney, Dry Pond (located on the hillside above Lake Bonney), and Canada stream that flows into Lake Fryxell (16 km from Lake Bonney). A limited number of nifH sequences were obtained (3 soil and 12 ice aggregate-associated nifH sequences) containing similar cyanobacterial clades as detected in our study (Fig. 4). Jungblut & Neilan (2010) also sequenced nifH genes from a cyanobacterial mat within a meltwater pond of the McMurdo Ice Shelf; however, in contrast to our study, all nifH transcript phylotypes were related to the Nostoc genus within the cyanobacteria (Jungblut & Neilan, 2010).

Delta-proteobacterial-related phylotypes dominated the gDNA nifH gene clone libraries (Table 3). nifH sequences of cultured delta-proteobacteria are typically from anaerobic sulfate reducers that group within cluster III of the four phylogenetic nifH clusters as proposed by Chien & Zinder (1996) and updated by Zehr et al. (2003a, b). These organisms also provide similar phylogenies when grouped according to 16S rRNA gene sequences (Zehr et al., 2003a, b). A large number of our delta-proteobacterial nifH sequences were related to Geobacter sp. (nifH subcluster IA; Duc et al., 2009). Geobacter species are able to fix N<sub>2</sub> under oxic conditions (Holmes et al., 2004), with some species capable of utilizing oxygen as an alternative electron acceptor (Methe et al., 2003; Lin et al., 2004). Therefore, cycles of aerobic and anaerobic conditions, as characterized by dry and wet periods of hyporheic zones and lake margins, may have a selective advantage for Geobacter species. This has also recently been suggested by Duc et al. (2009), who reported that 21% of nifH clones detected in the fore-field soils of a receding alpine glacier in Central Switzerland exposed to fluctuating periods of snowmelt and periods of drought were also related to Geobacter species. Nonphototrophicassociated N2 fixation rates in Antarctic habitats have also been detected in lake ice aggregates of the DV and mats of the McMurdo ice shelf under dark incubation conditions and incorporation of a photosystem II inhibitor (3-[3.4-dichlorophenyl]-1, 1-dimethyl urea; Olson et al., 1998; Paerl & Priscu, 1998; Fernández-Valiente et al., 2001; Jungblut & Neilan, 2010). Heterotrophic N<sub>2</sub> fixation activities were not detected in the ice shelf samples (Fernández-Valiente et al., 2001; Jungblut & Neilan, 2010), and heterotrophs accounted for approximately 10-30% of N<sub>2</sub> fixation activities detected within ice aggregate collected from a depth of 1.5 m in Lake Bonney of the DV (Olson et al., 1998; Paerl & Priscu, 1998). Many sulfate-reducing bacteria are known to contain nifH (Zehr et al., 1998) and have been shown to carry out a substantial portion of N<sub>2</sub> fixation in benthic marine environments (Capone, 1982; Steppe & Paerl, 2002; Bertics et al.,

2010). Similarly, in this study, sulfate reduction was linked with a large proportion (approximately 50%) of the N<sub>2</sub> fixation activities in the ML1-2 and MS1-3 transect sites (rates dropped from approximately 2.7 to 1.4 nmol N cm<sup>-2</sup> h<sup>-1</sup>) with high levels of sulfate reduction-linked N2 fixation also detected at other wetted DV soil sites (Fig. 3). Steppe & Paerl (2002) have also employed inhibition of sulfate reduction by the addition of MoO4 in a marine intertidal microbial mat and found that nighttime nitrogenase activity was inhibited by up to 64% with daytime rates remaining unaffected. The detection of sulfate-dependent N2 fixation and the identification of nifH associated with sulfate reducers (i.e. delta-proteobacterial group 2, Fig. 4) in the gDNA and the cDNA for ML1-2 may also indicate a similar link between the sulfur and nitrogen cycles in wetted Antarctic soils, suggesting heterotrophic diazotrophs are important contributors to N2 fixation.

Although the nitrogen cycle is poorly understood in soils of the DV (Barrett *et al.*, 2002), it is assumed that cyanobacteria are the major contributors to N<sub>2</sub> fixation owing to their conspicuous presence within microbial communities of wetted soils (Takacs-Vesbach *et al.*, 2010). For the first time, we report rates of N<sub>2</sub> fixation and identify *nifH* phylotypes and diversity within high-productivity soils of the DV. Collectively, our results suggest that both cyanobacteria and a diverse range of heterotrophic diazotrophs contribute to total N<sub>2</sub> fixation rates in wetted DV soils. However, further work is needed to define the dominance and niche differentiation between the autotrophic and heterotrophic N<sub>2</sub>-fixing populations within these most extreme environments.

# **Acknowledgements**

We would like to thank the staff at the United States Antarctic Program, as well as Antarctica New Zealand, and the Foundation for Research in Science and Technology, New Zealand, for logistical support while in the field. This research was supported by National Science Foundation Grants ANT 0739633 (to D.G.C.), ANT 0739640 (to E.J.C.), and ANT 0739648 and 0229836 (to S.C.C.).

# References

Adams BJ, Bardgett RD, Ayres E, Wall DH, Aislabie J, Bamforth S *et al.* (2006) Diversity and distribution of Victoria Land biota. *Soil Biol Biochem* **38**: 3003–3018.

Barrett JE, Virginia RA & Wall DH (2002) Trends in resin and KCl-extractable soil nitrogen across landscape gradients in Taylor Valley, Antarctica. *Ecosystems* 5: 289–299.

Barrett JE, Virginia RA, Lyons WB, McKnight DM, Priscu JC, Doran PT, Fountain AG, Wall DH & Moorhead DL

- (2007) Biogeochemical stoichiometry of Antarctic Dry Valley ecosystems. *J Geophys Res* **112**: 1–12.
- Bertics VJ, Sohm JA, Treude T, Chow CT, Capone DG, Fuhrman JA & Ziebis W (2010) Burrowing deeper into benthic nitrogen fixation: the impact of bioturbation on nitrogen fixation coupled to sulfate reduction. *Mar Ecol Prog Ser* **409**: 1–15.
- Bustin SA, Benes V, Garson JA, Hellemans J, Huggett J, Kubista M, Mueller R et al. (2009) The MIQE Guidelines: minimum Information for Publication of Quantitative Real-Time PCR Experiments. Clin Chem 55: 4.
- Capone DG (1982) Nitrogen fixation (acetylene reduction) by rhizosphere sediments of the eelgrass Zostera marina. Mar Ecol Prog Ser 10: 67–75.
- Capone DG (1993). Determination of nitrogenase activity in aquatic samples using the acetylene reduction procedure. Handbook of Methods in Aquatic Microbial Ecology (Kemp PF, Sherr BF, Sherr EB & Coles JJ, eds), pp. 621–631. CRC Press LLC, Boca Raton.
- Cary SC, McDonald IR, Barrett JE & Cowan DA (2010) On the rocks: the microbiology of Antarctic Dry Valley soils. *Nat Rev Microbiol* 8: 129–138.
- Charpy L, Palinska K, Casareto B, Langlade M, Suzuki Y, Abed R & Golubic S (2010) Dinitrogen-fixing cyanobacteria in microbial mats of two shallow coral reef ecosystems. *Microb Ecol* **59**: 174–186.
- Chien YT & Zinder SH (1996) Cloning, functional organization, transcript studies, and phylogenetic analysis of the complete nitrogenase structural genes (nifHDK2) and associated genes in the archaeon *Methanosarcina barkeri* 227. *J Bacteriol* 178: 143–148.
- Clarke KR & Gorley RN (2006) *PRIMER v6: User Manual*. Plymouth, UK: Plymouth Marine Laboratory.
- Cowan D, Russell N, Mamais A & Sheppard D (2002) Antarctic Dry Valley mineral soils contain unexpectedly high levels of microbial biomass. *Extremophiles* **6**: 431– 436.
- Danovaro R, Luna GM, Dell'Anno A & Pietrangeli B (2006) Comparison of two fingerprinting techniques, Terminal Restriction Fragment Length Polymorphism and Automated Ribosomal Intergenic Spacer Analysis, for determination of bacterial diversity in aquatic environments. Appl Environ Microbiol 72: 5982–5989.
- Duc L, Noll M, Meier B, Bürgmann H & Zeyer J (2009) High diversity of diazotrophs in the forefield of a receding alpine glacier. *Microb Ecol* 57: 179–190.
- Fernández-Valiente E, Quesada A, Howard-Williams C & Hawes I (2001) N<sub>2</sub>-Fixation in cyanobacterial mats from ponds on the McMurdo Ice Shelf, Antarctica. *Microb Ecol* **42**: 338–349.
- Fernández-Valiente E, Camacho A, Rochera C, Rico E, Vincent WF & Quesada A (2007) Community structure and physiological characterization of microbial mats in Byers Peninsula, Livingston Island (South Shetland Islands, Antarctica). *FEMS Microbiol Ecol* **59**: 377–385.
- Gooseff MN, McKnight DM, Runkel RL & Duff JH (2004) Denitrification and hydrologic transient storage in a glacial

- meltwater stream, McMurdo Dry Valleys, Antarctica. *Limnol Oceanogr* **49**: 1884–1895.
- Gordon DA, Priscu J & Giovannoni S (1998) Origin and phylogeny of microbes living in permanent Antarctic Lake ice. *Microb Ecol* 39: 197–202.
- Hewson I & Fuhrman JA (2006) Spatial and vertical biogeography of coral reef sediment bacterial and diazotroph communities. *Mar Ecol Prog Ser* **306**: 79–86.
- Holmes DE, Nevin KP & Lovley DR (2004) *In situ* expression of *nifD* in *Geobacteraceae* in subsurface sediments. *Appl Environ Microbiol* **70**: 7251–7259.
- Holm-Hansen O, Lorenzen CJ, Holmes RW & Strickland JDH (1965) Fluorometric determination of chlorophyll. *ICES J Mar Sci* **30**: 3–15.
- Hopkins DW, Sparrow AD, Elberling B, Gregorich EG, Novis PM, Greenfield LG & Tilston EL (2006a) Carbon, nitrogen and temperature controls on microbial activity in soils from an Antarctic dry valley. Soil Biol Biochem 38: 3130–3140.
- Hopkins DW, Sparrow AD, Novis PM, Gregorich EG, Elberling B & Greenfield LG (2006b) Controls on the distribution of productivity and organic resources in Antarctic dry valley soils. *Proc Biol Sci* 273: 2687–2695.
- Hopkins DW, Sparrow AD, Shillam LL, English LC, Dennis PG, Novis P, Elberling B, Gregorich EG & Greenfield LG (2008) Enzymatic activites and microbial communities in an Antarctic dry valley soil: reponses to C and N supplementation. *Soil Biol Biochem* **40**: 2130–2136.
- Izquierdo J & Nüsslein K (2006) Distribution of extensive nifH gene diversity across physical soil microenvironments. Microb Ecol 51: 441–452.
- Jungblut AD & Neilan BA (2010) NifH gene diversity and expression in a microbial mat community on the McMurdo Ice Shelf, Antarctica. Antarct Sci 22: 117–122.
- Kennedy AD (1993) Water as a limiting factor in the Antarctic terrestrial environment: a biogeographical synthesis. Arct Antarct Alp Res 25: 308–315.
- Lane DJ (1991). 16S/23S rRNA sequencing. Nucleic acid technology in bacterial systematics (Stackebrandt E & Goodfellow M, eds), pp 115–175. John Wiley and Sons, New York, NY.
- Lin WC, Coppi MV & Lovley DR (2004) Geobacter sulfurreducens can grow with oxygen as a terminal electron acceptor. Appl Environ Microbiol 70: 2525–2528.
- Liu X, Lindemann WC, Whitford WG & Steiner RL (2000) Microbial diversity and activity of disturbed soil in the northern Chihuahuan Desert. Biol Fertil Soils 32: 243–249.
- Maurice PA, McKnight DM, Leff L, Fulghum JE & Gooseff MN (2002) Direct observations of aluminosilicate weathering in the hyporheic zone of an Antarctic Dry Valley stream. Geochim Cosmochim Acta 66: 1335–1357.
- McKnight DM, Niyogi DK, Alger AS, Bomblies A, Conovitz PA & Tate CM (1999) Dry valley streams in Antarctica: ecosystems waiting for water. *Bioscience* **49**: 985–995.
- McKnight DM, Runkel RL, Tate CM, Duff JH & Moorhead DL (2004) Inorganic N and P dynamics of Antarctic glacial meltwater streams as controlled by hyporheic exchange and

- benthic autotrophic communities. JN Am Benthol Soc 23: 171–188
- McKnight DM, Tate CM, Andrews ED, Niyogi DK, Cozzetto K, Welch K, Lyons WB & Capone DG (2007) Reactivation of a cryptobiotic stream ecosystem in the McMurdo Dry Valleys, Antarctica: a long-term geomorphological experiment. *Geomorphology* **89**: 186–204.
- Methe BA, Nelson KE, Eisen JA, Paulsen IT, Nelson W, Heidelberg JF *et al.* (2003) Genome of *Geobacter sulfurreducens*: metal reduction in subsurface environments. *Science* **302**: 1967–1969.
- Niederberger TD, McDonald IR, Hacker AL, Soo RM, Barrett JE, Wall DH & Cary SC (2008) Microbial community composition in soils of Northern Victoria Land, Antarctica. *Environ Microbiol* **10**: 1713–1724.
- Novis PM, Whitehead D, Gregorich EG, Hunt JE, Sparrow AD, Hopkins DW, Elberling B & Greenfield LG (2007) Annual carbon fixation in terrestrial populations of *Nostoc commune* (Cyanobacteria) from an Antarctic dry valley is driven by temperature regime. *Glob Change Biol* 13: 1224–1237.
- Olson JB, Steppe TF, Litaker RW & Paerl HW (1998) N<sub>2</sub>-Fixing microbial consortia associated with the ice cover of Lake Bonney, Antarctica. *Microb Ecol* **36**: 231–238.
- Oremland RS & Capone DG (1988). Use of specific inhibitors in biogeochemistry and microbial ecology. Advances in Microbial Ecology (Marshall KC, ed.), pp. 285–383. Plenum Press, New York.
- Paerl HW & Priscu JC (1998) Microbial Phototrophic, heterotrophic, and diazotrophic activities associated with aggregates in the permanent ice cover of Lake Bonney, Antarctica. *Microb Ecol* **36**: 221–230.
- Pointing SB, Chan Y, Lacap DC, Lau MCY, Jurgens JA & Farrell RL (2009) Highly specialized microbial diversity in hyper-arid polar desert. *P Natl Acad Sci USA* **106**: 19964–19969.
- Priscu JC, Fritsen CH, Adams EE, Giovannoni SJ, Paerl HW, McKay CP, Doran PT, Gordon DA, Lanoil BD & Pinckney JL (1998) Perennial Antarctic Lake Ice: an oasis for life in a Polar desert. *Science* **280**: 2095–2098.
- Runkel RL, McKnight DM & Andrews ED (1998) Analysis of transient storage subject to unsteady flow: diel flow variation in an Antarctic stream. *JN Am Benthol Soc* 17: 143–154.
- Schloss PD & Handelsma J (2005) Introducing DOTUR, a computer program for defining operational taxonomic units and estimating species richness. *Appl Environ Microbiol* 71: 1501–1506.
- Smith J, Tow L, Stafford W, Cary C & Cowan D (2006) Bacterial diversity in three different Antarctic cold desert mineral soils. *Microb Ecol* 51: 413–421.
- Smith JS, Barrett JE, Tusnády G, Rejtö L & Cary SC (2010) Resolving environmental drivers in Antarctic soils. *Antarctic Sci* 22: 673–680.
- Steppe TF & Paerl HW (2002) Potential N<sub>2</sub> fixation by sulfatereducing bacteria in a marine intertidal microbial mat. *Aquat Microb Ecol* **28**: 1–12.

- Steppe TF & Paerl HW (2005) Nitrogenase activity and nifH expression in a marine intertidal microbial mat. Microb Ecol 49: 315–324.
- Takacs-Vesbach C, Zeglin L, Barrett JE, Gooseff MN & Priscu JC (2010). Chapter 8: Factors promoting microbial diversity in the McMurdo Dry Valleys, Antarctica. *Life in Antarctic deserts and other cold environments: Astrobiological Analogs* (Doran P, Lyons WB & McKnight DM, eds), pp. 221–257. Cambridge University Press, Cambridge.
- Thompson JD, Higgins DG & Gibson TJ (1994) CLUSTAL W: improving the sensitivity of progressive multiple sequence alignment through sequence weighting, position-specific gap penalties and weight matrix choice. *Nucleic Acids Res* 22: 4673–4680.
- Ueda T, Suga Y, Yahiro N & Matsuguchi T (1995) Remarkable N<sub>2</sub>-fixing bacterial diversity detected in rice roots by molecular evolutionary analysis of *nifH* gene sequences. *J Bacteriol* 177: 1414–1417.
- Vincent WF (1988). Microbial Ecosystems of Antarctica. Cambridge University Press, New York.
- Vincent WF (2000). Cyanobacterial dominance in the Polar Regions. *The Ecology of Cyanobacteria* (Whitton BA & Potts M, eds), pp. 321–340. Kluwer Academic Publishers, Dordrecht, The Netherlands.
- Vincent WF & Howard-Williams C (1986) Antarctic stream ecosystems: physiological ecology of a blue-green algal epilithon. *Freshwater Biol* **16**: 219–233.
- Vincent WF, Downes MT, Castenholtz RW & Howard-Williams C (1993) Community structure and pigment organisation of cyanobacteria-dominated microbial mats in Antarctica. Eur J Phycol 28: 213–221.
- Virginia RA & Wall DH (1999) How soils structure communities in the Antarctic dry Valleys. *Bioscience* **49**: 973–983.
- Wood SA, Rueckert A, Cowan DA & Cary SC (2008) Sources of edaphic cyanobacterial diversity in the Dry Valleys of Eastern Antarctica. ISME J 2: 308–320.
- Yeager CM, Kornosky JL, Housman DC, Grote EE, Belnap J & Kuske CR (2004) Diazotrophic community structure and function in two successional stages of biological soil crusts from the Colorado Plateau and Chihuahuan Desert. *Appl Environ Microbiol* **70**: 973–983.
- Young JPW (1992). Phylogenetic classification of nitrogenfixing organisms. *Biological Nitrogen Fixation* (Stacey G, Burris RH & Evans HJ, eds), pp. 43–86. Chapman and Hall, New York.
- Zani S, Mellon MT, Collier JL & Zehr JP (2000) Applied and Environmental Microbiology **66**: 3119–3124.
- Zeglin L, Sinsabaugh R, Barrett J, Gooseff M & Takacs-Vesbach C (2009) Landscape Distribution of Microbial Activity in the McMurdo Dry Valleys: linked Biotic Processes, Hydrology, and Geochemistry in a Cold Desert Ecosystem. *Ecosystems* 12: 562–573.
- Zehr JP & McReynolds LA (1989) Use of degenerate oligonucleotides for amplification of the *nifH* gene from the

- marine cyanobacterium *Trichodesmium thiebautii*. Appl Environ Microbiol **55**: 2522–2526.
- Zehr JP, Mellon M, Braun S, Litaker W, Steppe T & Paerl HW (1995) Diversity of heterotrophic nitrogen fixation genes in a marine cyanobacterial mat. Appl Environ Microbiol 61: 2527–2532.
- Zehr JP, Mellon MT & Zani S (1998) New nitrogen-fixing microorganisms detected in oligotrophic oceans by amplification of nitrogenase (*nifH*) Genes. *Appl Environ Microbiol* **64**: 3444–3450.
- Zehr JP, Jenkins BD, Short SM & Steward GF (2003a) Nitrogenase gene diversity and microbial community structure: a cross-system comparison. *Environ Microbiol* 5: 539–554.
- Zehr JP, Crumbliss LL, Church MJ, Omoregie EO & Jenkins BD (2003b) Nitrogenase genes in PCR and RT-PCR reagents and implications for studies of diverse functional genes. *Biotechniques* **35**: 995–1005.

# **Supporting Information**

Additional Supporting Information may be found in the online version of this article:

**Fig. S1.** Rarefaction curves (95% cut-off) for *nifH* gene clone libraries of BL2, NP2, MS1-3, ML1-2 and ML1-2 cDNA.

**Table S1.** Terminal-restriction fragments (TRFs) from each sample site utilized for PCA-based analyses.

**Table S2.** Percentage (total number of clones in parentheses) of *nifH* gene sequences in common between sites (95% sequence similarity).

Please note: Wiley-Blackwell is not responsible for the content or functionality of any supporting materials supplied by the authors. Any queries (other than missing material) should be directed to the corresponding author for the article.