# Application of pre-fishery abundance modelling and Bayesian hierarchical stock and recruitment analysis to the provision of precautionary catch advice for Irish salmon (Salmo salar L.) fisheries 

N. Ó Maoiléidigh, P. McGinnity, E. Prévost, E. C. E. Potter, P. Gargan, W. W. Crozier, P. Mills, and W. Roche


#### Abstract

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Ireland has one of the last remaining commercial salmon driftnet fisheries in the North Atlantic, with recent catches averaging 162000 salmon (1997-2003), approximately $20 \%$ of the total landings of salmon in the entire North Atlantic. Since 2001, the Irish commercial salmon fishery has been managed on the basis of Total Allowable Catch (TAC) in each of 17 salmon fishing districts. This has been made possible by applying a number of new and innovative techniques to the estimation of conservation limits (CLs) and prefishery abundance (PFA) for combined stocks in each district. Stock and recruitment parameters from well-monitored European rivers were "transported" to all Irish rivers using a Bayesian hierarchical stock and recruitment (BHSRA) model. This provided the posterior probability distributions of the model parameters and related reference points, including individual river CLs. District PFA and the number of spawners were estimated for a baseline period of 1997-2003, using district catch data, estimates of unreported catch, and exploitation rate. Harvest guidelines were established on the basis of surplus of spawning fish over the CL for the baseline period. In line with scientific advice, the commercial fishery has been reduced from 212000 fish in 2002 to 182000 in 2003. In 2004, a total catch (including the rod catch) of approximately 160000 wild salmon was advised.


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N. Ó Maoiléidigh and P. McGinnity: Marine Institute, Newport, Co. Mayo, Ireland. E. Prévost: INRA, UMR EQHC, 65 rue de St Brieuc, 35042 Rennes Cedex, France. E. C. E. Potter: CEFAS, Lowesoft Laboratory, Pakefield Road, Lowestoft, Suffolk NR33 0HT, England, UK. P. Gargan and W. Roche: Central Fisheries Board, Dublin 9, Ireland. W. W. Crozier: DARDNI, Agricultural and Environmental Sciences Division, Newforge Lane, Belfast BT9 5PX, Northern Ireland, UK. P. Mills: Compass Informatics Ltd., 18 Nassau Street, Dublin 2, Ireland. Correspondence to N. Ó Maoiléidigh: tel/fax: +35398 41107; email: niall.omaoileidigh@marine.ie.

## Introduction

Until 2002, the Irish fishery for salmon (Salmo salar) was managed by a combination of effort limitation and the application of technical conservation measures relating to size and type of fishing gear. While these measures regulate the efficiency of the fishery, they are not sensitive to the stock available, and allow the same level of fishing even when stocks are low. A Salmon Management Task Force
established in Ireland in 1996 (Anon., 1996) recommended a new rationale for management of salmon stocks. It was based on achieving "spawning escapement targets" for each specific stock, and maintaining stocks above conservation limits (CLs). The Task Force proposed the application of a Total Allowable Catch (TAC) to allow sufficient fish to spawn to meet the CL.

In 1998, the North Atlantic Salmon Conservation Organisation (NASCO, 1998) adopted the precautionary
approach to fisheries management (as outlined in FAO, 1995, 1996). Central to this was agreement that management measures should be aimed at maintaining all salmon stocks in the NASCO Convention Area above pre-agreed conservation limits.

The 17 salmon fishing districts in Ireland represent the smallest administrative entities for the management of commercial and recreational fishing, and for the provision of catch statistics. Their boundaries are statutorily defined and consist of the rivers, lakes, and estuaries, and an area of sea extending out from the land. In 1997, new legislation reduced the sea limit from 12 to 6 miles, and the fishing week from 5 days per week to 4 (daylight only). Salmon are taken mainly by surface driftnets in each of the district sea areas, although in some districts, draft nets are locally important (i.e. an estuarine or river seine-net held at one end on the shore, while the net is paid out by boat, the circle being completed by the boat returning to shore, entrapping any fish within the closing circle of net).

The homewater fisheries harvest two main components of the stock on their return to their natal rivers from their high-seas feeding grounds, i.e. 1 Sea Winter (1SW) salmon, generally referred to as "grilse", which return during summer, and 2 Sea Winter (2SW) salmon or "spring" salmon, which return during spring. In common with most other North Atlantic salmon stocks, the non-maturing component of the stock remains in the high seas, and is not exploited by the homewater fishery. The incidence of 2SW "summer salmon", and higher sea age groups is extremely low in Ireland. Similarly, the incidence of repeat spawners, from returns to monitored rivers and broodstock programmes, is low.

TACs can be set relative to stock abundance if a number of key values can be estimated, i.e. the number of fish available prior to the fishery taking place (i.e. the prefishery abundance, PFA), an estimate of the number of fish returning to spawn, and finally the number of fish required to spawn (the conservation limit). These values have been estimated for each of the 17 salmon fishing districts in Ireland, using PFA and pseudo stock and recruitment (PSR) models (Potter et al., 1998; Ó Maoiléidigh et al., 2001b, c; ICES, 2003).

More recently, the development of Bayesian hierarchical stock and recruitment analyses (BHSRA) for transporting stock and recruitment parameters from monitored rivers to rivers without stock and recruitment data (Crozier et al., 2003; Prévost et al., 2003) has allowed CLs to be established for individual rivers independently of catch data. The BHSRA relies on a measure of the production units available in each river. In the absence of a more refined and commonly available habitat variable, an intermediate, such as wetted area, can be used to quantify salmon production for widespread transport of stock and recruitment parameters from well-studied monitored rivers, provided the quantity of "non-productive" riverine habitat is in similar proportion from one river to the next. Accordingly,
a Geographical Information System (GIS) has been developed for the salmon rivers of Ireland (McGinnity et al., 2003), to provide an estimate of the size of the wetted area of each river in any given jurisdiction.

Here, we summarize the progress in management arrangements and fisheries for Irish salmon stocks, an important component of the overall North Atlantic salmon population. Several of the techniques used formed part of the work of the SALMODEL project initiated as an ECfunded Concerted Action (Crozier et al., 2003). The overall aim of SALMODEL was to "Advance the scientific basis upon which advice is given to managers of local, national, and international salmon fisheries, compatible with the precautionary approach, as adopted by the North Atlantic Salmon Conservation Organisation (NASCO) and within the requirement of sustainability". The application of these new and innovative techniques to the national management of a highly prized, but vulnerable, wild salmon resource is presented. New aggregated District and National CLs, generated from the BHSRA/wetted area approach, are compared with CLs generated from the previously applied PFA and PSR approach updated with 2003 catch, unreported catch, and data on the rate of exploitation. The status of Ireland's salmon stocks is illustrated relative to the attainment of this new CL, and the catch advice for each district is illustrated.

## Material and methods

## Estimation of pre-fishery abundance (PFA)

Catch records from commercial salmon dealers' registers are available for the period 1971-2000. Following implementation of a salmon carcass tagging and logbook scheme in 2001 (O Maoiléidigh et al., 2001a; Anon., 2004), the catch data derive from the logbook returns of commercial and recreational fishers.

The ICES model used to estimate the pre-fishery abundance of salmon from countries in the NEAC area employs a simple run-reconstruction approach similar to that described by Potter and Dunkley (1993) and Rago et al. (1993). Following Potter et al. (1998), the model takes the catch in numbers of 1 SW salmon in each district, then raises it to account for estimates of non-reported catches and exploitation rates. The last two parameters are generated from a comprehensive national coded wire tagging and tag recovery programme (Browne, 1982; Ó Maoiléidigh et al., 1996). As unreported catch cannot be estimated precisely, minimum and maximum values likely to encompass the true values are provided, based on local information or inspection. Similarly, exploitation rate data for individual districts are based on the estimated range of values for the monitored rivers in those districts or the next nearest districts. These ranges are used to delimit uniform distributions for the parameters in a Monte Carlo simulation. An example of the input data required for the
national CL is shown in Table 1; a similar table is required for each of the 17 salmon fishing districts. The simulation, which uses the software package "Crystal Ball" (Risk assessment software add-in for Microsoft Excel ${ }^{\mathrm{TM}}$, from Decisioneering Inc, Decisioneering UK, Ltd, 1996) is run in Microsoft Excel ${ }^{\mathrm{TM}}$ to generate estimated distributions of the PFA values by simulating 1000 runs of the model. The PFA values are raised to take account of the natural mortality between 1 January in the first sea winter (which is the date they recruit to the first fishery, i.e. Faroes) to provide an estimate of pre-fishery abundance (PFA) or recruitment (Potter et al., 1998). Subsequently, the PFA is discounted to

Table 1. Input data for Irish pre-fishery abundance (PFA) analysis using Monte Carlo simulation. Estimated catch (1SW salmon only) derives from licenced salmon dealers' registers up to and including 2000, and from mandatory commercial fishing logbooks to 2003. Unreported catch and exploitation rate derive from the National Coded Wire Tagging and Tag Recovery programme.

| Year | $\begin{gathered} \begin{array}{c} \text { Catch } \\ \text { (numbers) } \end{array} \\ \hline 1 \mathrm{SW} \end{gathered}$ | Percentage unreported |  | Exploitation rate 1SW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Minimum | Maximum | Minimum | Maximum |
| 1971 | 417428 | 30 | 45 | 56 | 74 |
| 1972 | 449160 | 30 | 45 | 56 | 74 |
| 1973 | 460665 | 30 | 45 | 56 | 74 |
| 1974 | 561323 | 30 | 45 | 56 | 74 |
| 1975 | 616249 | 30 | 45 | 56 | 74 |
| 1976 | 420508 | 30 | 45 | 56 | 74 |
| 1977 | 368579 | 30 | 45 | 56 | 74 |
| 1978 | 324350 | 30 | 45 | 56 | 74 |
| 1979 | 289539 | 30 | 45 | 56 | 74 |
| 1980 | 237561 | 30 | 45 | 56 | 74 |
| 1981 | 157713 | 30 | 45 | 51 | 68 |
| 1982 | 277527 | 30 | 45 | 60 | 79 |
| 1983 | 463602 | 30 | 45 | 56 | 75 |
| 1984 | 243152 | 30 | 45 | 54 | 72 |
| 1985 | 456437 | 30 | 45 | 63 | 84 |
| 1986 | 509992 | 30 | 45 | 59 | 79 |
| 1987 | 344066 | 20 | 40 | 61 | 82 |
| 1988 | 416652 | 20 | 40 | 56 | 75 |
| 1989 | 316536 | 20 | 40 | 60 | 80 |
| 1990 | 183589 | 20 | 40 | 57 | 76 |
| 1991 | 116923 | 20 | 40 | 47 | 62 |
| 1992 | 180868 | 20 | 40 | 60 | 80 |
| 1993 | 152577 | 15 | 35 | 51 | 68 |
| 1994 | 235935 | 15 | 35 | 55 | 74 |
| 1995 | 233313 | 15 | 35 | 62 | 83 |
| 1996 | 202582 | 15 | 35 | 53 | 71 |
| 1997 | 152808 | 10 | 20 | 54 | 71 |
| 1998 | 162054 | 10 | 20 | 53 | 70 |
| 1999 | 145336 | 10 | 20 | 52 | 70 |
| 2000 | 180823 | 10 | 20 | 49 | 65 |
| 2001 | 234682 | 5 | 10 | 58 | 77 |
| 2002 | 198633 | 5 | 10 | 48 | 64 |
| 2003 | 166151 | 5 | 10 | 48 | 64 |

account for natural mortality ( $0.03 \%$ monthly instantaneous rate of natural mortality), to estimate the total returns back to homewaters. The spawning population is then estimated by subtracting the catch from the returns.

## Estimation of conservation limits (CLs)

## Pseudo stock-recruitment (PSR) model

Following Potter et al. (1998), estimates of spawning stocks are derived as model outputs from the information on catches, unreported catch, and exploitation rate. However, stock-recruitment relationships cannot be derived directly from these data, because the spawners in year " $n$ " contribute to the 1 SW recruitment in year " $\mathrm{n}+3$ " to year " $\mathrm{n}+5$ ", depending on the relative proportions of $1-3$ -year-old smolts that can be produced. Thus, spawners in year " $n$ " may produce: 1-year-old smolts in year " $n+2$ ", which generate 1 SW recruits in year " $n+3$ "; 2-year-old smolts in year " $\mathrm{n}+3$ ", which produce 1 SW recruits in year " $n+4$ "; and so forth.

The egg deposition in year " n " is assumed to contribute to the recruitment in year " $\mathrm{n}+3$ " to " $\mathrm{n}+5$ " in proportion to the numbers of smolts produced of ages $1-3$. Therefore, the number of "lagged eggs" that contribute to the recruitment of maturing 1SW fish in each year can be estimated. The lagged egg estimates provide a measure of the relative spawning level that gave rise to the recruitment figures expressed above as PFA. These data can then be plotted to provide a "pseudo" stock-recruitment relationship, and a number of reference points can be derived. It is not possible to estimate stock size at maximum sustainable yield ( $\mathrm{S}_{\mathrm{MSY}}$ ) from this relationship without making further assumptions about marine survival. As a result, a method for setting biological reference points from noisy stock-recruitment relationships was developed by Potter and Nicholson (2001). The model assumes that there is a critical stock level below which recruitment decreases linearly towards zero stock and recruitment, and above which recruitment is constant. The position of the critical stock level is determined by searching for the value that minimizes the residual sum of squares. Potter and Nicholson (2001) applied this approach to similar pseudo stock-recruitment relationships. An example of the output for the national stock based on the input data in Table 1 is shown in Figure 1.

## Bayesian hierarchal stock and recruitment analysis (BHSRA)/wetted area

The analysis of stock and recruitment (SR) data is the most widely used approach for deriving biological reference points (BRPs) for salmon populations (Prévost et al., 2001). While the conservation limits generated from the PSR model use stock and recruitment data for each district, they are "pseudo" because they relate to geographic entities (i.e. the number of fish returning to that district) rather than true biological stocks. They are further complicated by the


Figure 1. Example of a pseudo stock and recruitment curve for the total Irish stock. Recruits are expressed in numbers of fish available prior to the first fishery, i.e. Faroes. This date is taken as 1 January in the first sea winter. Spawning stock is expressed in number of eggs. The position of the critical stock level is determined by searching for the value that minimizes the residual sum of squares (approximately 4 billion eggs, equivalent to 171600 1SW spawners).
mixed stock nature of these district fisheries. Browne et al. (1994) and Ó Maoiléidigh et al. (1994) showed that a significant number of tagged fish from specific rivers are caught in districts other than the district of origin.

Prévost et al. (2003) applied Bayesian hierarchical modelling of stock-recruitment (SR) relationships to estimate BRPs for European Atlantic salmon stocks. The structure of the hierarchical SR model developed distinguishes two nested levels of randomness, within-river and between-rivers. This is an extension of the classical Ricker modelling approach, where the parameters of the Ricker function are assumed to be different between rivers, but drawn from a common probability distribution depending on two primary covariates, river size, and river latitude. The Bayesian analysis of this hierarchical model has been developed with a set of 13 stock and recruitment data series from monitored salmon rivers located in the Northeast Atlantic. The model yields a set of predicted SR parameters for new rivers with no SR data, provided information is available on the size of the river (in this case wetted area is used), and its latitude. Posterior distributions are estimated by means of MCMC sampling (Gibbs algorithm), as implemented by the Winbugs software (Spiegelhalter et al., 2000). Details of the model specification and its Bayesian treatment are given in Prévost et al. (2003).
The latitude value used in the analysis is the river catchment area midpoint, and the size is quantified as the riverine wetted area accessible to salmon. The wetted area is computed from statistically combined parameters, the length of upstream river, upstream catchment area, stream order, and local channel gradient, captured by aerial photography and extracted within the GIS platform (McGinnity et al., 2003).

Given this latitude and wetted area information, the approach described in Prévost et al. (2003) was used to estimate new district CLs, defined as the sum of riverspecific CLs for each fishery district.

## Derivation of precautionary catch advice

The catch data used in the models is modified to account for 2SW and older salmon, and hatchery-reared fish released as smolts. There is little overlap between the run timing of the main two components of the Irish stock, i.e. the main 1SW salmon run is from June to August, whereas 2SW (spring) salmon enter rivers during spring. As the commercial fishery operates in a restricted time period (June and July), it exploits predominantly 1 SW salmon. The exploitation rate on larger Multi-Sea Winter salmon (MSW) or "spring" fish is low, approximately $7.5 \%$ on average for the period 1980-1988, and declining (ICES, 2003), and legislation introduced in 1997 has eliminated commercial fishing until 12 May. As no further scale analyses have been carried out since 1988 , a value of $7 \%$ has been used as the proportion of 2 SW salmon in the commercial catches annually since that time. Although there may be some small contribution to subsequent 1 SW stocks from the progeny of older salmon, for the purposes of providing catch advice, 2 SW and older fish (including repeat spawners) are excluded from the catch data used in the models, principally because such stocks are not exposed to a significant commercial fishery, and represent only a relatively small proportion of the total population. Similarly, hatchery-reared fish released as smolts are not considered to make a contribution to spawning populations, and potential catches of these fish have been estimated and removed from the catch data. The
output from the models provides the average (1997-2003) number of spawners and the total catch of adult wild 1 SW salmon.

The next step is to convert the conservation limits in eggs $\mathrm{m}^{-2}$ from the BHSRA/wetted area analysis to total egg requirement for each river in Ireland by multiplying by the total wetted area accessible to salmon. Subsequently, the values for egg deposition are converted to adults, and then corrected for 1SW fish only in the manner outlined below.

If the total number of adults to meet CL is X , then the number of eggs will be

$$
\begin{aligned}
& \left(\text { Prop }_{1 \text { SW }} \times \text { Prop }_{1 \text { SW Female }} \times \text { Eggs }_{1 \text { SW }} \times \mathrm{X}\right) \\
& \quad+\left(\text { Prop }_{2 \text { SW }} \times \text { Prop }_{2 \text { SW Female }} \times \text { Eggs }_{2 \text { SW }} \times \mathrm{X}\right)
\end{aligned}
$$

where Prop $_{1 \text { SW }}$ and Prop $_{2 \text { SW }}$ are the proportions of 1 SW and 2SW fish ( 0.93 and 0.07 , respectively, based on age distribution analyses from 1980 to 1988), and Prop $_{1 \text { SW Female }}$ and Prop $_{2 \text { SW Female }}$ are the proportions of 1SW female and 1SW male fish ( 0.6 and 0.4 , respectively), and the proportions of 2SW female and 2SW male fish ( 0.85 and 0.15 , respectively). These proportions are broadly based on observations of returning wild salmon in several rivers monitored in Ireland (e.g. River Burrishoole percentage 1SW females $1997-2002=59.9 \%$, percentage 2SW females for the same period $=83 \%$ ). Eggs ${ }_{1 \text { SW }}$ and Eggs $_{2 \text { SW }}$ are the average fecundities of 1 SW and 2SW fish (3400 and 8000 eggs, respectively). Again, these values are based broadly on monitored river data where broodstock are stripped, and eggs (including residual eggs) are counted.

The total number of adults, $\mathrm{X}=\mathrm{CL} / 2373$, and therefore the CL consists of $(0.93 \times X) 1 \mathrm{SW}$ and $(0.07 \times \mathrm{X}) 2 \mathrm{SW}$ salmon.

Once estimates of average spawners, average catch, and district CL are produced, precautionary catch advice is formulated by applying harvest guidelines, as follows. Where the average number of spawners is higher than the CL, an exploitable surplus exists. If this surplus is more than or equal to the average catch for the baseline period, this average catch can be taken in the next season. In following a precautionary approach, increases over this average are not permitted even if the surplus is higher, because of the mixed stock nature of the fishery, and the desire to protect more vulnerable stocks in other areas. Where the average number of spawners is less than the CL, the average catch must be reduced by this number of fish in the next season. Where there is no exploitable surplus or where the total returns before exploitation are less than the CL, there should be no catch, or the fishery should be severely restricted while stock rebuilding takes place.

## Results

The results showed 173 salmon rivers in Ireland located between 51.6 and $55.3^{\circ} \mathrm{N}$. They vary in size from 3727 to
$8795447 \mathrm{~m}^{2}$ of riverine wetted area accessible to salmon (median $182949 \mathrm{~m}^{2}$ ). There is wide overlap in the size of the Irish rivers and the size range of the monitored rivers used as a learning sample by Prévost et al. (2003), i.e. $10 \%$ of the Irish rivers are smaller than the smallest monitored river, but none are bigger than the largest one. The Irish rivers are grouped into 17 salmon fishing districts, plus that part of the River Foyle within the Republic of Ireland. The number of rivers in each fishery district varies from 1 to 30 .

Owing to the lognormal structure of the hierarchical SR model used, the posterior predictive distributions of the CLs are best examined on a log-scale (Figure 2). The distribution location (e.g. the median) of the egg deposition rates of the fishery district CLs vary little around that of the national CL, because of the narrow latitudinal range within Ireland. Indeed, Prévost et al. (2003) show that it is the latitude covariate that governs the variations in the location of the CL posterior predictive distributions. There are large variations in the precision of the individual district posterior predictive CLs. For instance, in the Drogheda district, which consists of just one river, the posterior predictive distribution ranges from approximately 0.5 to 16 eggs $\mathrm{m}^{-2}$. In districts where several rivers are aggregated, the CL predictions are more precise, e.g. the Kerry district ( 30 rivers) ranges from approximately 1.5 to $7 \mathrm{eggs} \mathrm{m}^{-2}$. The variance reduction effect gained from the aggregation of several rivers under a regional entity is more pronounced when the number of rivers increases. This explains why the CL rate of egg deposition at the national level is more precisely estimated than that of any individual fishery district. The relative size of the rivers within a fishery region also has an effect on the precision of the estimates. The CL of the Lismore fishery district, which has 7 rivers with a single large one accounting for more than $75 \%$ of the wetted area accessible to salmon in the district, is estimated with a similar level of precision as the Drogheda fishery district.

The $90 \%$ probability interval of the posterior predictive distributions of CLs generally encompass the point estimate CLs derived from the PFA/PSR approach previously used for providing catch advice in Ireland (Figure 2). However the PFA/PSR CLs are overdispersed compared with their corresponding posterior predictions using the BHSRA/ wetted area approach: only 5 of $17(29 \%)$ are located in the inter-quartile interval, and 12 of 17 (71\%) within the $80 \%$ probability interval. There is also a tendency of the PFA/ PSR estimates to be greater than the estimates derived from the BHSRA/wetted area values approach. Indeed, 10 of 17 (59\%) of the catch-based CL estimates are located in the upper half of their corresponding posterior distribution, and 6 of $17(35 \%)$ are in the upper quartile. The noticeable exception is the Dublin fishery district, where the previous estimate based on the PFS/PSR model was significantly underestimated. The national CL derived from the PFA/ PSR model results in a mean value of approximately 7 eggs $\mathrm{m}^{-2}$, and is located in the upper part of the posterior predictive distribution close to the 90 th percentile. This


Figure 2. Posterior predictive distributions of the egg deposition rate per $\mathrm{m}^{2}$ of riverine wetted area accessible to salmon, corresponding to the conservation limits of the fishery districts, and of Ireland as a whole. Each box plot displays on a log-scale the 5th, 10th, 25 th, 50 th, 75th, 90th, and 95 th percentiles. Black dots represent the currently used conservation limits for management advice in Ireland for the fisheries district, and at ICES for the whole country. The Dublin fishery district conservation limit does not appear on the graph because it is lower than the lower bound of the $y$-axis.
compares with the BHSRA/wetted area median value of approximately 4 eggs $\mathrm{m}^{-2}$.

The status of the district stocks relative to their attainment of mean post-predicted BHSRA/wetted area CLs in 2003 is shown in Figure 3. Of the 17 fisheries districts in Ireland, 6 meet their conservation limits, 6 are $>50 \%$ of their CL, and the remaining districts fall as low as $15 \%$ of their CL. Nationally, the attainment of CL for 1SW salmon is just over $80 \%$.

The precautionary catch advice is given in Table 2. This advice is predicated on wild fish only (i.e. estimated returns from hatchery-released smolts have been removed). It also relates to the total removal of fish by all means, and is not restricted to commercial fisheries. There are five districts, mainly located on the east and south coasts, where the conservation limit would not have been met even in the absence of a commercial fishery. Therefore, there is no catch indicated in the precautionary catch column. In seven of the districts, reductions in the average catch in 2004 are indicated, because these districts were below their conservation limits on average for the period prior to this. The remaining districts are meeting or exceeding their
conservation limits. In this instance, the average catch is advised for 2004, recognizing the fact that the fishery is a mixed stock one, taking fish from districts that are below their conservation limit. The status of these districts will be assessed on an ongoing basis, and the advice will change in line with any significant and consistent improvement in stock size.

## Discussion

Ideally, river-specific stock and recruitment analysis would be the most accurate way to determine river-specific conservation limits. However, given that river-specific stock and recruitment studies are resource-intensive and take a long time to cover several generations and a wide range of stock levels, the BHSRA/wetted area method represents the most feasible method of deriving individual river CLs for the foreseeable future (Prévost et al., 2001).

The two methods of estimating conservation limits are fundamentally different in approach and independent of each other. The approach described by Prévost et al. (2003)


Figure 3. Average attainment of the mean posterior predicted BHSRA/wetted area conservation limit (CL) from 1997 to 2003, for each Irish salmon fishing district.
uses a training set of 13 rivers, with annual fish population size data to estimate the stock and recruitment parameters. The method provides probability distributions of the number of eggs that can be accommodated, based on the limitations imposed by the physical attributes of each river. In contrast, the PFA/PSR models use catch data to derive an aggregated stock-recruitment relationship, without explicit reference to the physical limitation of the size of the rivers comprising the aggregate (district) being considered, and assuming that the fish belong to spawning populations originating in that district. Despite the two different
approaches used, the national CL based on the PFA/PSR approach ( 1716041 SW salmon) is not greatly different from the equivalent value using the mean of the postpredicted BHSRA/wetted area output (197576 1SW salmon). This tends to support the contention that the PFA/PSR models are robust means of estimating a national conservation limit, because all spawning stocks are included.

There was a variable degree of correspondence between the comparable estimates for each individual district. However, in just one case (the Dublin district) was the

Table 2. Precautionary catch advice for 2004 based on average attainment of conservation limits in each Irish salmon fishing district from 1997 to 2003.

|  |  | BHSRA/wetted <br> area conservation <br> limit | Average <br> number of <br> spawners | Estimated average <br> wild catch (including <br> rod catch) | Precautionary catch <br> of wild 1SW <br> salmon in 2004 |
| :--- | :--- | :---: | :---: | :---: | :---: |
| Region | District | Dundalk | 5037 | 2468 | 1478 |

estimate generated by the PFA/PSR model far outside the posterior predicted $90 \%$ probability range of the BHSRA/ wetted area output.

Even using best available information, the wide range of posterior predicted distributions of CLs associated with individual districts illustrates the uncertainty that managers are faced with, given that it is the mean CL that has been adopted to date. In applying precautionary catch advice, it is therefore recognized that this may not always afford adequate protection to stocks, and that fisheries that could possibly exploit at a higher rate are being restricted. However, the process itself is seen as more appropriate to the present fishery and stock levels. Development of the catch advice in a risk framework is also anticipated in the near future, and it will be possible to address some of the uncertainty issues as stocks and fisheries are monitored.

Other differences between comparative estimates at the district level could be accounted for by the following: (i) the mixed stock nature of district fisheries, and the problems this poses to the PFA/PSR models that use catch as the primary data input, i.e. the CL egg deposition rate requirement is overestimated if the district catches many fish from other districts without significant loss of its own fish to other district fisheries (possibly in Lismore, Ballina, and Letterkenny), and underestimated if the district loses a lot of fish to other districts without gaining fish from other districts (possibly in Dundalk, Drogheda, Ballinakill, and Sligo); (ii) the accuracy of the wetted area estimate, which is based on a regression model with its associated uncertainty; (iii) the degree to which water quality (or any other factor influencing the productive capacity of rivers) is compromised relative to other Irish rivers, and the 13 rivers in the training set. This will lead to over- or underestimates by the BHSRA/wetted area approach, if water quality is more or less (respectively) compromised than the training set rivers. Deterioration in water quality is a serious issue in a number of districts in east and southeast Ireland (i.e. Dundalk, Dublin, Wexford, and Waterford). The differences in outputs between the models may be an indication or measure of the reduction in carrying capacity or productive potential of the rivers in these districts, as a result of water quality problems; (iv) the PFA/PSR models used for district catch advice do not include the 2SW and MSW (or repeat spawner) fish in the CL (i.e. no input data for this component is entered in the model), because the main fisheries under quota are grilse fisheries, and information on the relative catch of MSW stocks at a district level is poor. This could affect some of the individual districts where 2 SW or MSW fish constitute a larger proportion of the catch (notably Wexford and Lismore), and corrections may apply to these areas in future; (v) similarly, the index rivers used to provide exploitation rate data for the PFA/PSR model may not be representative of all districts; (vi) the PFA/PSR ICES model is based on a 30 -year catch series, and has not been corrected for non-stationarity. Crozier et al. (2003) reported
a consistent drop in productivity for 6 rivers in the Northeast Atlantic around the mid-1980s. Similarly, ICES (2003) identified a potential "phase-shift" in salmon production that occurs around 1990 (for adults) for rivers and stocks in both the Northeast Atlantic and in North America. The BHSRA/wetted area approach presented here only takes account of stock and recruitment parameters from the mid-1980s.

These factors could act discretely or cumulatively, depending on the circumstances and the dynamics of individual river populations. However, the derivation of CL probability distributions by the BHSRA/wetted area approach is an improvement on the point estimates of district CLs obtained from the PFA/PSR catch-based models, because it removes the uncertainty associated with the mixed stock nature of district fisheries. It also allows for a more in-depth appraisal of the underlying biology of individual stocks in relation to the productive capacity of the river producing them. Moreover, these river CLs can potentially be refined with more information on the physical characteristics of the catchments (compromised water quality, gradient, etc) to a higher level of precision.

While district BHSRA/wetted area CLs are preferred over the estimates derived from PFA/PSR models, it is likely that, for the foreseeable future, the PFA model will continue to be used to generate a retrospective estimate of exploitable surplus, and to assess the attainment of these new district CLs. This situation may improve with information from automatic fish counters, which could be used as indices of performance for all other rivers in each district, and to derive forward-running predictive models for pre-fishery abundance. Fish counters would also provide additional information on stock structure and run timing, which will refine many of the parameters currently used in models and in the provision of catch advice to managers.
Improvements in the collection of salmon catch statistics (mandatory carcass tagging, logbooks), and advances in analytical techniques have permitted the derivation of precautionary catch advice which is now being used to establish TACs for each district, and to apportion the surplus over spawning requirements to the various fishing engines (driftnets, draft nets, other nets, rods). In the short period of implementation from 2002 to 2004, district TACs have been applied as far as possible with reference to those districts requiring the most significant reductions in catch. However, reductions in catch alone will not in many instances result in an instant response in spawning stocks. Clearly, significant rebuilding will be required in some areas, while surpluses in spawning requirement in other districts will need to be examined in relation to the extent of mixing of stocks originating from other districts in each district's catch. Most importantly, the combined techniques have allowed the implementation of a new national management plan that provides the basis for maintaining Irish salmon stocks above conservation limits, or to rebuild deficient stocks.

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